

GLYPHOSATE (N-(PHOSPHONOMETHYL)GLYCINE) CONCENTRATIONS IN WATER COURSES – SYSTEMATIC REVIEW AND SCIENTOMETRIC ANALYSIS

CONCENTRAÇÕES DE GLIFOSATO (N-(FOSFONOMETIL)GLICINA) EM CURSOS DE ÁGUA – REVISÃO SISTEMÁTICA E ANÁLISE CIENTOMÉTRICA

CONCENTRACIONES DE GLIFOSATO (N-(FOSFONOMETIL)GLICINA) EN CURSOS DE AGUA – REVISIÓN SISTEMÁTICA Y ANÁLISIS CIENCIOMÉTRICO

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Sandria Ferreira Cavassani¹, Karla da Silva Malaquias², Michelle Nauara Gomes do Nascimento³, Isadora Barboza Silva⁴, Sandra Aparecida Benite-Ribeiro⁵

ABSTRACT

Glyphosate, which degrades into aminomethylphosphonic acid (AMPA), is the most widely used active ingredient in herbicides worldwide. Both compounds can enter aquatic systems through surface runoff, leaching, spray drift, and irrigation, leading to water contamination and subsequent incorporation into the food chain. This study aimed to perform a systematic review and scientometric analysis of research published between 2015 and 2025 on glyphosate and AMPA concentrations in surface and groundwater, and to compare geographically detected concentrations with national regulatory thresholds. A systematic review was conducted following the PRISMA protocol, complemented by scientometric analysis. Literature searches were performed in the Web of Science, PubMed, ScienceDirect, and SciELO databases. A total of 127 articles reporting glyphosate and AMPA concentrations in surface and groundwater were selected. The countries contributing the largest number of studies were Argentina, Brazil, Canada, the United States, Mexico, and Italy. Scientometric analysis revealed that these nations not only dominate research output but also constitute the most influential co-citation networks, with the most frequently cited study originating from the United States. The highest concentration reported was in Brazil (8,700 µg/L), which is 133 times above the Brazilian regulatory limit (65 µg/L). Statistical analyses further showed that glyphosate concentrations vary significantly by geographic region, with notable differences between Europe and North America. Glyphosate concentrations frequently exceed national maximum permissible limits, even in countries with stringent legislation such as those in Europe, where values surpassed the legal threshold of 0.1 µg/L at multiple sites. These findings underscore the widespread nature of glyphosate contamination and highlight the need for stronger monitoring and regulatory enforcement.

E-mail: sandriacavassani@discente.ufj.edu.br

E-mail: michelle.nascimento@ufj.edu.br

E-mail: sandrabenite@ufj.edu.br

¹ Master in Applied Health Sciences. Universidade Federal de Jataí, Goiás, Brazil.

² Doctor of Chemistry.Universidade Federal de Jataí, Goiás, Brazil. E-mail: ksmalaquias@ufj.edu.br

³ Post-doctorate in Chemistry. Universidade Federal de Jataí, Goiás, Brazil.

⁴ Master in Applied Health Sciences. Universidade Federal de Jataí, Goiás, Brazil.

E-mail: isadorabarboza@discente.ufj.edu.br

⁵ Doctor of Biological Sciences. Universidade Federal de Jataí, Goiás, Brazil.



Keywords: Surface Water. Groundwater. Glyphosate. Contamination. Systematic Review. Scientometrics.

RESUMO

O glifosato, que se degrada em ácido aminometilfosfônico (AMPA), é o ingrediente ativo mais amplamente utilizado em herbicidas em todo o mundo. Ambos os compostos podem entrar em sistemas aquáticos por meio de escoamento superficial, lixiviação, deriva de pulverização e irrigação, levando à contaminação da água e subsequente incorporação na cadeia alimentar. Este estudo teve como objetivo realizar uma revisão sistemática e análise cienciométrica de pesquisas publicadas entre 2015 e 2025 sobre concentrações de glifosato e AMPA em águas superficiais e subterrâneas, e comparar concentrações detectadas geograficamente com limites regulatórios nacionais. Uma revisão sistemática foi conduzida seguindo o protocolo PRISMA, complementada por análise cienciométrica. Buscas bibliográficas foram realizadas nas bases de dados Web of Science, PubMed, ScienceDirect e SciELO. Um total de 127 artigos relatando concentrações de glifosato e AMPA em águas superficiais e subterrâneas foram selecionados. Os países que contribuíram com o maior número de estudos foram Argentina, Brasil, Canadá, Estados Unidos, México e Itália. A análise cienciométrica revelou que essas nações não apenas dominam a produção científica, mas também constituem as redes de cocitação mais influentes, com o estudo mais frequentemente citado originário dos Estados Unidos. A maior concentração relatada foi no Brasil (8.700 μg/L), o que é 133 vezes acima do limite regulatório brasileiro (65 μg/L). As análises estatísticas mostraram ainda que as concentrações de glifosato variam significativamente por região geográfica, com diferenças notáveis entre a Europa e a América do Norte. As concentrações de glifosato frequentemente excedem os limites máximos permitidos nacionais, mesmo em países com legislação rigorosa como os da Europa, onde os valores ultrapassaram o limite legal de 0,1 µg/L em vários locais. Essas descobertas ressaltam a natureza generalizada da contaminação por glifosato e destacam a necessidade de monitoramento e fiscalização regulatória mais rigorosos.

Palavras-chave: Águas Superficiais. Águas Subterrâneas. Glifosato. Contaminação. Revisão Sistemática. Cienciometria.

RESUMEN

El glifosato, que se degrada en ácido aminometilfosfónico (AMPA), es el ingrediente activo más utilizado en herbicidas a nivel mundial. Ambos compuestos pueden ingresar a los sistemas acuáticos a través de la escorrentía superficial, la lixiviación, la deriva de pulverización y el riego, lo que lleva a la contaminación del agua y la posterior incorporación a la cadena alimentaria. Este estudio tuvo como objetivo realizar una revisión sistemática y un análisis cienciométrico de las investigaciones publicadas entre 2015 y 2025 sobre las concentraciones de glifosato y AMPA en aguas superficiales y subterráneas, y comparar las concentraciones detectadas geográficamente con los umbrales regulatorios nacionales. Se realizó una revisión sistemática siguiendo el protocolo PRISMA, complementada con un análisis cienciométrico. Se realizaron búsquedas bibliográficas en las bases de datos Web of Science, PubMed, ScienceDirect y SciELO. Se seleccionó un total de 127 artículos que informaban sobre las concentraciones de glifosato y AMPA en aguas superficiales y subterráneas. Los países que aportaron el mayor número de estudios fueron Argentina, Brasil, Canadá, Estados Unidos, México e Italia. El análisis cienciométrico reveló que estas naciones no solo dominan la producción de investigación, sino que también constituyen las redes de cocitación más influyentes, y el estudio más citado proviene de Estados Unidos. La



concentración más alta reportada se registró en Brasil (8700 μ g/L), 133 veces superior al límite regulatorio brasileño (65 μ g/L). Los análisis estadísticos mostraron además que las concentraciones de glifosato varían significativamente según la región geográfica, con diferencias notables entre Europa y Norteamérica. Las concentraciones de glifosato frecuentemente exceden los límites máximos permisibles nacionales, incluso en países con una legislación estricta como los europeos, donde los valores superaron el umbral legal de 0,1 μ g/L en múltiples sitios. Estos hallazgos subrayan la naturaleza generalizada de la contaminación por glifosato y resaltan la necesidad de un monitoreo y una aplicación regulatoria más rigurosos.

Palabras clave: Aguas Superficiales. Aguas Subterráneas. Glifosato. Contaminación. Revisión Sistemática. Cienciometría.



1 INTRODUCTION

Glyphosate (N-(phosphonomethyl)glycine) is one of the most extensively used herbicides worldwide, particularly in agricultural systems where it is applied for weed control and pre-harvest grain desiccation (Bianco *et al.*, 2023; Connolly; Coggins; Koch, 2020). Applied mainly via spraying, glyphosate undergoes degradation largely through microbial activity and plant metabolism, generating aminomethylphosphonic acid (AMPA) as its primary metabolite (Delmonico *et al.*, 2014). Since their introduction to the market in the 1970s (Bianco *et al.*, 2023), glyphosate-based herbicides (GBHs) have become the most widely used class of herbicides globally. Their primary mechanism of action in plants and certain bacteria involves inhibition of 5-enolpyruvylshikimate-3-phosphate synthase (EPSPS), a key enzyme in the shikimate pathway (Mesnage *et al.*, 2015). Because this pathway is absent in vertebrates, early assumptions suggested that glyphosate posed minimal risk to human health (Marino *et al.*, 2021; Mesnage *et al.*, 2015).

More recent evidence, however, has increasingly challenged this premise, pointing to potential indirect effects associated with long-term exposure. Although humans are not the intended targets, exposure can occur through ingestion of contaminated water (Ruiz-Toledo *et al.*, 2014), occupational contact (Bianco *et al.*, 2023), and dietary intake (Marino *et al.*, 2021). Multiple pathways of surface and groundwater contamination have also been documented, including heavy rainfall (Rodrigues *et al.*, 2019), spray drift (Aranha *et al.*, 2023), agricultural runoff (Aranha *et al.*, 2023), and leaching (Gunarathna *et al.*, 2018).

A broad spectrum of adverse health outcomes has been associated with human exposure to glyphosate and its metabolites, including endocrine disruption (Curwin *et al.*, 2006), oxidative stress, carcinogenicity (Bukowska *et al.*, 2022), chromosomal damage (De Souza *et al.*, 2020), type 2 diabetes mellitus (Pandey; Dabhade; Kumarasamy, 2019), inhibition of acetylcholinesterase (AChE) activity (Larsen *et al.*, 2016), as well as cardiometabolic and hepatic disorders (Bukowska *et al.*, 2022; Curwin *et al.*, 2006; De Souza *et al.*, 2020; Eskenazi *et al.*, 2023; Larsen *et al.*, 2016; Pandey; Dabhade; Kumarasamy, 2019). In 2015, the International Agency for Research on Cancer (IARC) classified glyphosate as a probable human carcinogen (Group 2A). However, in 2016, both the Food and Agriculture Organization of the United Nations (FAO) and the World Health Organization (WHO) re-evaluated this designation and ultimately removed glyphosate from Group 2A, citing insufficient conclusive evidence (Mendonça *et al.*, 2020; Van Bruggen *et al.*, 2018).



Divergent perspectives regarding human health risks, combined with the continued growth in glyphosate usage, have led several countries — including Italy, Canada, Spain, and Portugal — to impose bans on this herbicide (Brovini *et al.*, 2021). Nevertheless, recent studies have reported elevated concentrations of glyphosate and its primary metabolite AMPA in water bodies even within countries enforcing restrictions, such as Italy (Feltracco *et al.*, 2022b; Gomarasca *et al.*, 2024) and Canada (Brown; Farenhorst, 2024; Edge *et al.*, 2023; Montiel-León *et al.*, 2019). These findings highlight both the environmental persistence of glyphosate and the limitations of regulatory measures in effectively mitigating aquatic contamination.

The contamination of water resources by glyphosate and AMPA has thus emerged as a globally relevant issue, requiring systematic assessment to elucidate its extent as well as research trends. In this context, scientometric and spatial analyses serve as strategic approaches for mapping the state of scientific production, identifying geographic and thematic gaps, and guiding future research agendas (de Castilhos Ghisi *et al.*, 2020). Accordingly, the present review pursues two main objectives: I. To analyze spatial patterns of glyphosate and AMPA contamination by: a) mapping the geographic distribution of concentrations through manual data extraction and compiling comparative tables of minimum and maximum values by country; and b) comparing detected concentrations with the prevailing national regulatory thresholds. II. To evaluate global scientific production on glyphosate and AMPA contamination in aquatic environments through scientometric analysis, focusing on: a) the evolution of publications and citations; b) the most frequently cited authors; and c) the most influential publications.

2 MATERIALS AND METHODS

a) Scope of Analysis:

The systematic review was conducted in accordance with the PRISMA protocol (Swartz, 2011), We assembled peer-reviewed articles reporting glyphosate and/or AMPA concentrations in groundwater (artesian and deep wells) and surface waters (rivers, lakes, and streams), published between 2015 and 2025 in English or Portuguese. Searches were performed in SciELO, ScienceDirect, Web of Science, and PubMed databases using predefined Boolean strings in both English and Portuguese. Examples of queries in English included Glyphosate OR AMPA AND surface water and Glyphosate OR AMPA AND



groundwater. Equivalent terms were applied in Portuguese (e.g., Glifosato OR AMPA AND água subterrânea).

b) Eligibility Criteria:

Studies were included if they: (i) reported glyphosate and/or AMPA concentrations in groundwater or surface waters; (ii) described the study environment and analytical methods; (iii) were peer-reviewed journal articles published in English or Portuguese; and (iv) fell within the 2015–2025 timeframe. Exclusion criteria comprised: review/meta-analysis articles, retracted papers, methodological validation studies, experimental model studies (e.g., animals, microcosms), studies lacking concentration data (only frequency), or reporting concentrations solely in graphical format.

c) Data Extraction and Systematic Review:

After duplicate removal (RStudio, v.4.3.3), titles, abstracts, and full texts were screened against inclusion criteria. Eligible studies were catalogued in Microsoft Excel (2010, Office 14). Extracted information included: (i) authorship and year; (ii) country, region, or waterbody; (iii) analytical method used; (iv) evaluated compounds; and (v) minimum and maximum concentrations. Data extraction served two purposes: (a) mapping global distributions of reported concentrations, and (b) comparing observed levels with national regulatory thresholds.

d) Statistical Analysis

Statistical analyses were conducted in IBM SPSS Statistics (v.31.0.0.0). Data were stratified by major categories identified in the literature. According to distribution, parametric or non-parametric tests were applied. The Kruskal–Wallis test with Tukey post-hoc was used for group comparisons. Significance was set at p < 0.05. Outliers were screened using the boxplot method but retained in analyses.

e) Scientometric Analysis:

Selected bibliographic records (full text and cited references) were exported in plain text format ("download_XXX.txt") and processed individually for each database. Scientometric analysis was performed using CiteSpace (v.6.3.R1), a Java-based platform for knowledge domain visualization (Zhou *et al.*, 2023). CiteSpace was employed to build:



Collaboration networks (authors, institutions, countries); Journal co-citation maps; Keyword co-occurrence networks; Temporal evolution of research hotspots. Node size represented frequency or citation count, while node color indicated temporal citation activity (Lin *et al.*, 2022). Time slicing was set from January 2015 to December 2025, selecting the top 20–50 most cited records per slice. The software shows the results in map form, the nodes indicate the search items and links between nodes describe quotes or mutual references between nodes. Rings are formed with different colors representing each node; the colors represent the chronological order of occurrence of links and items (de Castilhos Ghisi *et al.*, 2020). The blue color represents older articles, while the newer articles are orange, containing an outer purple ring that indicates good centrality (or yellow circles). Centrality is defined in the **(1)**.

Centrality (node
$$i$$
) = $\sum i \neq j \neq k \frac{Pjk(i)}{Pjk}$ (1)

In (1), pjk represents the number of shortest paths between node j and node k, and pjk (i) is the number of these paths passing through node i. The analysis of bibliographic coupling was also used. The indicator used to evaluate the grouping is the coefficient of silhouette (Li et al., 2017), defined in the Eq. (2).

$$S(i) \begin{cases} 1 - a(i)/b(i), & \text{if } a(i) < b(i), \\ 0, & \text{if } a(i) = b(i), \\ b(i)/a(i) - 1, & \text{if } a(i) > b(i) \end{cases}$$
 (2)

In equation (2), DS, a(i) corresponds to the average distance of i from all other data within the same cluster, b(i) s the smallest average distance of i from any other cluster of which i is not a member, and s(i) is the coefficient of silhouette and has interval [-1,1], given in Eq. (3).

$$-1 \le s(i) \le 1 \tag{3}$$

Additionally, QGIS (v.3.34.3) was used to generate cartographic visualizations of the global distribution of studies, enabling the integration of scientometric and spatial perspectives.



3 RESULTS AND DISCUSSION

The review revealed that the highest glyphosate concentrations were detected in South American and North American countries, with Brazil reporting the maximum concentration — $8,700 \, \mu g/L$, equivalent to 133 times above the national regulatory limit (65 $\, \mu g/L$). This result underscores a marked discrepancy between environmental contamination levels and legal thresholds, suggesting shortcomings in both monitoring practices and regulatory enforcement of agricultural herbicide use.

The scientometric analysis demonstrated a progressive growth in publications addressing glyphosate and AMPA dissipation across environmental matrices, particularly in the last decade. A parallel trend was observed for studies examining toxicological risks to human health (Zhang *et al.*, 2022), including endocrine disruption, oxidative stress (Bukowska *et al.*, 2022), and chronic disease outcomes, reflecting the increasing societal and scientific concern surrounding these compounds.

The initial bibliographic search retrieved 87,036 records across the selected databases: 50,127 from Web of Science, 35,907 from ScienceDirect, 973 from PubMed, and 29 from SciELO (Fig. 1). After duplicate removal, 16,210 unique entries remained. Non-relevant publications — such as studies focused solely on environmental risk assessments, toxicological experiments using animal models, or unrelated subject areas — were excluded. Ultimately, 127 articles met the inclusion criteria and formed the analytical dataset. This substantial reduction from initial retrieval to final inclusion highlights the fragmentation of the scientific literature, with a large portion addressing glyphosate-related themes without directly quantifying residues in aquatic environments. The final dataset, therefore, represents a highly specific but globally relevant corpus of research, suitable for comparative analysis of geographic contamination patterns and aligned evaluation with national regulatory standards.



Figure 1

Systematic review diagram of glyphosate concentrations in waters worldwide. Figure adapted from the methodology PRISMA. (Swartz, 2011)

Identification

Studies identified in the survey databases:

Web of Science: 50,127 Science Direct: 35,907

PubMed: 973 Scielo: 29 Total: N= 87.036

Screening and eligibility

Studies after removal of duplicates and retracted: N=16.211

Studies excluded after reading the title and abstract: N=16.082

No access in full:

N = 02

Inclusion

Studies included in the review:

3.1 DESCRIPTIVE/SYSTEMATIC DATA ANALYSIS

3.1.1 Occurrence of Gly and AMPA in waters worldwide

The 127 studies included in this review were distributed across 35 countries, with the largest research outputs originating from Argentina (23 articles), Brazil (17), Canada (12), the United States (7), Mexico (6), and Italy (6) (Fig 2; Table 1). Argentina, Brazil, and the United States together account for 82% of global soybean production, a crop that relies heavily on glyphosate for weed management and pre-harvest desiccation. These countries are also characterized by comparatively less stringent regulatory frameworks, which, along with intensive agricultural practices, help explain both their extensive glyphosate usage and higher research activity in this area (Brovini *et al.*, 2021; Connolly; Coggins; Koch, 2020). Brazil represents a central case: in 2023, national records reported sales of 253,302 tons of glyphosate and its derivatives (IBAMA, 2023). This intensive use has been paralleled by a



marked increase in studies evaluating glyphosate and AMPA concentrations in surface and groundwater, reflecting growing environmental and regulatory concerns within the country.

In the Asia-Pacific region, countries such as China and India are among the world's largest consumers of glyphosate (Richmond, 2018). Yet, our review identified only four studies in China, one in Malaysia, and two in Thailand (**Fig. 2**), highlighting a clear mismatch between glyphosate consumption and environmental monitoring efforts.

In contrast, Europe emerged as the second-largest hub of scientific production on glyphosate contamination in waters, following the Americas (Fig. 2; Fig 3). Within the European Union, glyphosate usage has fluctuated considerably due to regulatory uncertainties and debates regarding its risk categorization (Richmond, 2018). These controversies have led several member states to adopt precautionary restrictions on glyphosate and glyphosate-based herbicides (Marino et al., 2021), which may partly account for the comparatively high number of monitoring studies conducted across Europe.

Table 1Maximum and minimum concentrations found of glyphosate and the metabolite AMPA in surface and groundwater

Nº	Min-Max	Min-Max	Country	Site	Guideline	References
	(AMPA)	(Gly)	6 11 ÁC:	C (.	(μg/L)	(11 1 2040)
1		<lod-0,42< th=""><th>South África</th><th>Surface water</th><th>-</th><th>(Horn et al. 2019)</th></lod-0,42<>	South África	Surface water	-	(Horn et al. 2019)
	N.M.	μg/L				
2	0,03-0,21 μg/L	0,03-0,90 μg/L	Germany	Surface water	0,1	(Tauchnitz et al.,
-						2020)
3	62,1-2633ng/L	11,2-106ng/L	Germany	Surface water	0,1	(Wirth et al. 2021)
4	0,2-6,5 μg/L	0,6-21,2 μg/L	Argentina	Suface water;	240	(Demonte et al.,
				Groundwater		2018)
5	0,75-42 μg/L	0,50-97 μg/L	Argentina	Surface water	240	(Alonso et al.,
						2018)
6	0,44-69,90	0,48-3,01 μg/L	Argentina	Groundwater	240	(Giacobone et al.,
	μg/L					2023)
7	0,7-49,4 μg/L	0,2-167,4 μg/L	Argentina	Surface water	240	(Lutri <i>et al.,</i> 2020)
8	0,77-0,90 μg/L	1,25-4,52 μg/L	Argentina	Surface water	240	(Berman et al.,
						2018)
9	N.D0,5 μg/L	N.D.	Argentina	Surface water	240	(Lupi <i>et al.</i> , 2015)
10	0,100-2,600	0,100-20,50	Argentina	Groundwater	240	(Aparicio;
	μg/L	μg/L				Gerónimo, 2024)
11	2,3-233 µg/L	4,54-111	Argentina	Suface	240	(Vera-Candioti et
		μg/L	•	water;		<i>al.</i> , 2021b)
				Groundwater		,
12	0,2-5,1 μg/L	0,1-35 µg/L	Argentina	Surface	240	(Mas <i>et al.</i> , 2020)
				water		
13	0,32-2,00	0,78-2,90	Argentina	Surface	240	(Pérez et al.,
-	μg/L	μg/L		water		2017)



14	N.D5 μg/L	N.D19 μg/L	Argentina	Surface water	240	(Van Opstal <i>et al.</i> , 2023)
15	N.D. 4,8 μg/L	N.D125 μg/L	Argentina	Surface water	240	(Bonansea <i>et al.</i> , 2017)
16	N.D.	0,4-1,2 μg/L	Argentina	Surface water	240	(Ronco <i>et al.</i> , 2016)
17	0-0,10µg/L	0-1,7µg/L	Argentina	Surface water	240	(Jesabel Perez et al., 2017)
18	0 0,10µg/L	0 1,7 µg/L	7 ti goritina	Surface	240	(Mac Loughlin <i>et</i>
.0	0,2-4,5µg/L	0,2-17µg/L	Argentina	water	210	al., 2020)
19	N.D	N.D	, g =	Surface	240	(Peluso <i>et al.</i> ,
	2,13µg/L	7,62µg/L	Argentina	water		2022)
20	, -1 J	, , , , , , , , , , , , , , , , , , ,	<u> </u>	Suface	240	(Primost <i>et al.</i> ,
	0,53-	0,73-		water;	-	2017)
	1,90µg/L	1,80µg/L	Argentina	Groundwater		,
21	, 10	Not		Surface	240	(Avigliano;
	N.M.	available-		water		Schenone,
		1600µg/L	Argentina			2015a) [′]
22				Surface	240	(Andrade et al.,
	<0,6-4µg/L	<0,6-7µg/L	Argentina	water		2021)
23	·		<u> </u>	Surface	240	(Peluso <i>et al.</i> ,
	N.D4µg/L	7-12µg/L	Argentina	water		2020)
24	N.D	7-12μg/L N.D		Surface	240	(Peluso et al.,
	10,3µg/L	22,4µg/L	Argentina	water		2023)
25				Suface	240	(Okada <i>et al.</i> ,
				water;		2018)
	0,1-3,7µg/L	0,1-8,5µg/L	Argentina	Groundwater		
26	Not	Not		Surface	240	(Mayora <i>et al.</i> ,
	available-	available-		water		2024)
	4μg/L	3µg/L	Argentina			
27	1,3-10 µg/L	1,1-14,2 μg/L	Australia	Surface	1000	(Okada <i>et al.</i> ,
				water		2020)
28	0,55-2,42	1,95-2,96	Australia	Surface	1000	(Okada <i>et al.</i> ,
	μg/L	μg/L	A 4 !!	water	4000	2019)
29	∠0 E0 20"	<0.0.070····/	Australia	Surface	1000	(Campbell <i>et al.</i> ,
	<0,50-32µg/L	<0,2-370µg/L		water	0.4	2025)
30	7.2.202~//	0.70.1E2~/	Poleium	Surface	0,1	(Quaglia <i>et al.</i> ,
24	7,2-303µg/L	0,78-153µg/L	Belgium	water	0.1	(Tang et al
31	0.22 5 0	0.64.6.4	Poleium	Surface	0,1	(Tang <i>et al.</i> ,
22	0,23-5,8µg/L	0,64-6,1µg/L	Belgium	water	65	(Dovidouskas et
32	N.D.	N.D13,9	Brazil	Surface	65	(Dovidauskas <i>et</i> <i>al.</i> , 2022)
33	N.D.	μg/L 0.6078-11.33	Brazil	water Surface	65	(Correia;
33	IN.D.		ומבוו	water	00	Carbonari; Velini,
		μg/L		walei		2020)
34	0,65-1,9 μg/L	1,5-9,7 µg/L	Brazil	Suface	65 e 500	(Pires <i>et al.</i> ,
U-T	0,00-1,0 μg/L	1,0 0,7 µg/L	DIGZII	water;	00 0 000	2020)
				Groundwater		2020)
35	N.D.	0,5-8,7 mg/L	Brazil	Surface	65	(Lima <i>et al.</i> , 2023)
30	11.0.	5,5 5,7 mg/L	DIGE	water	50	(Linia ot al., 2020)
36	N.D.	11,6-45,5	Brazil	Surface	65	(Damiani <i>et al.</i> ,
J.		µg/L	2.32.	water	20	2023)
37	N.M.	0,035-0,77	Brazil	Surface	65	(Feliciano <i>et al.</i> ,
٠.		mg/L		water		2025)
38		<u> </u>	Brazil	Suface	65 e 500	(Pires <i>et al.</i> ,
-	0,0026-	0,0062-		water;		2023b)
	0,2751µg/L	1,5868µg/L		Groundwater		- /
	, , ,	, , ,				



39			Brazil	Suface	65 e 500	(Camiccia et al.,
	N.M.	0,001-		water;		2022)
		0,802µg/L		Groundwater		·
40	N.M.	N.M0,72	Brazil	Surface	65	(Prezilius et al.,
		µmol L		water		2022)
41	23,46-	51,88-	Brazil	Surface	65	(Gomes <i>et al.</i> ,
71			Diazii	water	00	2022)
40	41,25µg/L	117,07µg/L	D		0.5	
42	0,50-	0,31-	Brazil	Surface	65	(Mendonca <i>et al.</i> ,
	1,40µg/L	1,65µg/L		water		2020)
43	N.M.	12,58-	Brazil	Groundwater	500	(Nunes <i>et al.</i> ,
		65,11µg/L				2024)
44	N.D.	N.D.	Brazil	Surface		(Didone et al.,
				water		2021)
45		1.90 x	Brazil	Surface	65	(Ohse et al.,
.•	N.M.	10^16-5.27 x	D. a.L.ii	water	00	2024)
	I V.IVI.	10^10_5:27 X 10^10 pmol/L		Water		2024)
40	N.D.		Deseil	Curfoso	C.F.	/Dimenstrate
46	N.D.	N.D.	Brazil	Surface	65	(Pimenta <i>et al.</i> ,
				water		2020)
47			Brazil	Suface	65 e 500	(de Matos <i>et al.</i> ,
				water;		2023)
	26-234µg/L	27-371µg/L		Groundwater		•
48	Not	Not available		Surface	65	(Gabardo et al.
	available-		Brazil and	water.		2024)
	0,85µg/L		Paraguay	water.		2024)
40		4F C CC2		Craumduratan	200	() /on Ctomony ont
49	5-698 ng/L	15,6-663	Canada	Groundwater	280	(Van Stempvoort
		ng/L				et al., 2016)
50	N.D.	N.D17 μg/L	Canada	Surface	65	(Edge <i>et al.</i> ,
				water		2023)
51	17-3034 ng/L	25-3380 ng/L	Canada	Surface	65	(Struger et al.
	· ·	· ·		water		2015)
52	33-656 ng/L	2,8-3.000	Canada	Surface	65	(Montiel-León et
-	00 000 1.g/L	ng/L	Janaaa	water	00	al., 2019)
53		TIG/L	Canada	Surface	65	(Raby <i>et al.</i> ,
33	1 75 14ua/l	NIM 10ua/I	Carraua		05	
	1,75-14µg/L	N.M19µg/L	0 1 .	water	0.5	2022)
54	N.M	N.M	Canada	Surface	65	(Donald et al.
	2900ng/L	17645ng/L		water		2018)
55	N.D.	N.D.	Canada	Surface	65	(Sanford; Prosser,
				water		2020)
56		0,07-	Canada	Surface	65	(Picard et al.,
	0,76-9,2ng/L	0,60ng/L		water		2021)
57	<lod< th=""><th><lod< th=""><th>Canada</th><th>Surface</th><th>65</th><th>(ljzerman <i>et al.</i>,</th></lod<></th></lod<>	<lod< th=""><th>Canada</th><th>Surface</th><th>65</th><th>(ljzerman <i>et al.</i>,</th></lod<>	Canada	Surface	65	(ljzerman <i>et al.</i> ,
٠.	LOD	-205	Gariada	water	00	2024)
58	153-		Canada		65	
50		07.4005://	Carraua	Surface	03	(Feng <i>et al.</i> ,
	5836ng/L	87-4095ng/L		water		2024)
59		0,057-	Canada	Surface	65	(Brown;
	0,11-3,1µg/L	0,53µg/L		water		Farenhorst, 2024)
60	Not	Not	Canada	Surface	65	(Feng <i>et al.</i> ,
	available-	available-		water		2025)
	55,7ng/L	483ng/L				•
61	N.M.	N.D.		Surface	0,1	(Wang <i>et al.</i> ,
٠.			China	water	٥, .	2016)
62	<0,05-	<0,05-	Offilia	Sup; subt.	0 1	(Geng <i>et al.</i> ,
02	•		China	Sup, Subt.	0,1	
	10,31µg/L	32,49µg/L	China	0	2.4	2021)
63	N.M.	N.D.		Groundwater	0,1	(Jing <i>et al.</i> , 2021)
			China			
64		0,41-		Surface	0,1	(Tan <i>et al.</i> , 2024)
	2,5-37,1µg/L	37,1µg/L	China	water		,



66 N.M. 204,5-441,5µg/L 441,5µg/L µg/L Colombia vater Surface water - (Alvarez Bayona et al., 2022) 67 N.D. 0,12-6,24 Spain Surface water 0,1 (Melendez-Pastor) 68 10,2- 4,8ng/L Spain Suface water; 0,1 (Lopez-Vazquez et al., 2023) 69 0,05- 0,184µg/L 0,03- 0,228µg/L Spain Surface 0,1 (Yusa et al., 2021) 70 20-9500 ng/L 40-7900 ng/L United States Surface water 65 (Bradley et al., 2021) 71 N.M34,100 N.M91,700 United ng/L Surface states 65 (Byres et al., 2023) 72 N.M1700 N.M420 United United Surface water 65 (Battaglin et al., 2023) 73 United Surface water 65 (Battaglin et al., 2023) 74 N.M5,6µg/L N.M8,1µg/L States water 2021) 75 N.M. Quilled Surface water 65 (Melalie et al., 2021) 76 N.M. Quilled Surface wate	65	N.D.	N.D.	Groundwater - (Bahamón-		(Bahamón-Pinzón	
Fig. Surface Colombia Colombia Surface Colombia Colombia Surface Colombia Colombi		11.5.	14.5.	Colombia	Croundwater		
N.D. 0,12-6,24 Spain Surface water 0,1 (Melendez-Pastor) et al., 2021)	66	N.M.	204,5-		Surface	-	
Box							
Spain Suface 0,1 (Lopez-Vazquez et al., 2023)	67	N.D.		Spain		0,1	
1505.6- 10,2- 4.8 mg/L 3027, 5mg/L Groundwater			μg/L				
4,8/mg/L 3027/5/mg/L Groundwater 0.1 (Yusa et al., 2021) 70 0,055- 0,23- 0,228 μg/L surface water 0.2 1 (Yusa et al., 2021) 70 20-9500 ng/L 40-7900 ng/L United Surface water 65 (Bradley et al., 2027) 71 N.M34,100 N.M91,700 ng/L United Surface water 65 (Byers et al., 2023) 72 N.M1700 ng/L N.M30 United Surface water Surface 5 (Byers et al., 2023) 73 ng/L United Surface water 5 (Byers et al., 2023) 73 United Surface water 2020) (States water 2021) 74 N.M5,6µg/L N.M8,1µg/L States water 2020) 75 N.M. N.M8,1µg/L States water 2020) 75 N.M. United Surface water 65 (Mahier et al., 2017) 76 N.M. States water 2020) 75 N.M. States water 2019) 76 N.M. States water Surface water, 2017) </th <th>68</th> <th></th> <th></th> <th>Spain</th> <th></th> <th>0,1</th> <th></th>	68			Spain		0,1	
69 0.055- 0.184μg/L 2021) 0.03- 0.228μg/L 40-7900 ng/L 100-1000 ng/L 1							et al., 2023)
O.184µg/L 0.228µg/L water 2021) 70 20-9500 ng/L 40-7900 ng/L United States Surface 65 (Bradley et al., 2017) 71 N.M34,100 N.M91,700 United ng/L States water 2025) 72 N.M1700 N.M420 United states water 65 (Battaglin et al., 2023) 73 United ng/L States water 2023) 74 United N.M5,6µg/L N.M8,1µg/L States water 2021) 75 N.M. United Surface 65 (Mediale et al., 2020) 75 N.M. United Surface 65 (Melale et al., 2020) 75 N.M. States water 2020) 76 N.M. States water 2017) 76 N.M. 60-300ng/L States water 2017) 77 2,94(Switzerl Al., 400 g/L 8,40(Croatia) Europe* Surface 0,1 (Navarro et al., 2024) 79 <t< th=""><th></th><th></th><th></th><th>0</th><th></th><th>0.4</th><th>()/1 -1</th></t<>				0		0.4	()/1 -1
To 20-9500 ng/L 40-7900 ng/L United States Stat	69	•		Spain		0,1	
States Water 2017	70			l lucitor al		C.F.	
71 N.M34,100 ng/L N.M91,700 ng/L United States water Surface water 65 (Byers et al., 2025) 72 N.M1700 ng/L N.M420 ng/L United Surface water 65 (Battaglin et al., 2023) 73 0,50-2,5μg/L 1,6-13μg/L States water 65 (Ulrich; Ferguson, 2021) 74 United N.M5,6μg/L N.M8,1μg/L States water 2020) 75 N.M. United Surface water 65 (Mahler et al., 2017) 76 N.M. United Surface water 65 (Mahler et al., 2017) 76 N.M. States water 2019) 77 2,94(Switzerl and)-988 -3499 ng/L ng/L (Czech (Italy) Europe* Surface water 0,1 (Navarro et al., 2024) 78 1,4-3μg/L (A. 2004) 6.5-7,3μg/L (B. 2004) Europe* water 2018) 79 0,01-0,6 μg/L (A. 2004) 0,1-1,7 μg/L (B. 2004) France (B. 2004) Surface (B. 2004) 0,1 (B. 2004) 80 0,1 (A. 2004) France (B. 2004) Surface (B. 2004) 0,1 (B. 2004) 81 <th>70</th> <th>20-9500 ng/L</th> <th>40-7900 ng/L</th> <th></th> <th></th> <th>00</th> <th></th>	70	20-9500 ng/L	40-7900 ng/L			00	
Total	74	N.M. 24.400	N.M. 04 700			G.F.	
N.M4700	71					03	
Table	72					65	,
Table Tab	12					03	
1,6-13μg/L 1,	73	Hg/∟	TIG/L			65	
N.M 5,6μg/L N.M 8,1μg/L States Surface water	13	0.50-2.5ug/l	1 6-13ug/l			03	
N.M 5,6μg/L N.M 8,1μg/L States Water C2020 T5 N.M. C2-27,8μg/L United Surface Surface G5 (Mahler et al., 2017) T6 N.M. States Water C2020 N.M. States Surface G5 (Welch et al., 2017) T7 2,94(Switzerl 8,40(Croatia) and)-988 -3499 ng/L (Italy) Republic Republic Republic T8 1,4-3μg/L 6,5-7,3μg/L Europe* Surface Water C2024) T9 0,01-0,6 μg/L 0,1-1,7 μg/L France Surface Water C2021) T8 1,4-3μg/L 386,9μg/L France Surface O,1 (LeGrancq et al., 2021) T8 367 300 France Surface O,1 (LeGrancq et al., 2021) T8 367 300 France Surface O,1 (Reoyo-Prats et al., 2017) T8 3,66- 0,40 France Surface O,1 (Reoyo-Prats et al., 2017) T8 3,66- 0,40 France Surface O,1 (Merdy et al., 2017) T8 3,66- 0,40 France Surface O,1 (Merdy et al., 2017) T8 3,66- 0,40 France Surface O,1 (Merdy et al., 2017) T8 3,66- 0,40 France Surface O,1 (Merdy et al., 2025) T8 Netherland Surface O,1 (Desmet et al., 2025) T8 N.D. N.D. Netherland Surface O,1 (Geerdink et al., 2026) T8 N.D. N.D. Netherland Surface O,1 (Geerdink et al., 2024) T8 Surface O,1 (Szekacs; Moertl; Darvas, 2015) T8 N.M. 125-455ng/L Hungary Groundwater C4 C2023 T8 N.D. N.D1,32 Iran Surface O,1 (Szekacs; Moertl; Darvas, 2015) T8 N.D. N.D1,32 Iran Surface O,1 (Szekacs; Moertl; Darvas, 2015) T8 Surface O,1 (Samarghandi et al., 2022) T8 N.D. N.D1,32 Iran Surface O,1 (Samarghandi et al., 2023) T8 Surface O,1	74	0,00 2,0µg/L	1,0 10µg/L			65	,
N.M.		N.M5 6ug/l	N.M8 1ua/l			00	
N.M. States Water 2017 (Welch et al., 2019)	75		11 0, 1 µg/ L			65	
N.M. States Suface water; 2019 (Welch et al., 2019)	. •		<2-27.8ua/L			00	
N.M. 60-300ng/L States Water; Groundwater 2019	76		, - 1 J			65 e 700	
Company Comp		N.M.					
77 2,94(Switzerl and)-988 ng/L (Croatia) and)-988 ng/L (Italy) Europe* water Surface water 0,1 (Navarro et al., 2024) 78 Republic) Surface water 0,1 (Ginebreda et al., 2018) 79 0,01-0,6 μg/L 0,1-1,7 μg/			60-300ng/L				,
Telepholic Te	77	2,94(Switzerl		Europe*		0,1	(Navarro et al.,
Republic T8		and)-988	-3499 ng/L		water		2024)
78 1,4-3μg/L 6,5-7,3μg/L Europe* Surface water 0,1 (Ginebreda et al., 2018) 79 0,01-0,6 μg/L 0,1-1,7 μg/L France Surface water 0,1 (Le Cor et al., 2021) 80 0,1- France Surface water 0,1 (Lefrancq et al., 2017) 81 367- 300- 1500ng/L France water Surface water 0,1 (Reoyo-Prats et al., 2017) 82 3,66- 0,40- 5 rance water France Surface water 0,1 (Merdy et al., 2017) 83 Netherland s water Surface water 0,1 (Desmet et al., 2025) 84 111- y9μg/L Netherland water Surface water 0,1 (Geerdink et al., 2020) 85 N.D. N.D. N.D. Netherland surface Groundwater 0,1 (Broers et al., 2022) 86 Surface water 0,1 (Toth et al., 2022) 9,05-3μg/L 0,2-2μg/L Hungary water Water Darvas, 2015) 87 N.M. 125-455ng/L Hungary Hungary water Groundwater - (Pakzad et al., 2023) 89		ng/L (Czech	(Italy)				
1,4-3μg/L 6,5-7,3μg/L Europe* water 2018 79 0,01-0,6 μg/L 0,1-1,7 μg/L France Surface water 2021 80 0,1- France Surface 0,1 (Le Cor et al., 2021) 81 367- 300- France Surface 0,1 (Reoyo-Prats et al., 2017) 82 3,66- 0,40- France Surface 0,1 (Merdy et al., 2017) 83 Surface 0,1 (Merdy et al., 2017) 84 367- 300- France Surface 0,1 (Merdy et al., 2017) 85 3,66- 0,40- France Surface 0,1 (Merdy et al., 2015) 83 Netherland Surface 0,1 (Desmet et al., 2016) 0,02-130μg/L <0,02-12μg/L Netherland Surface 0,1 (Geerdink et al., 2016) 0,02-130μg/L <0,02-12μg/L S water 2020) 85 N.D. N.D. Netherland Groundwater 0,1 (Groundwater al., 2024) 86 Surface 0,1 (Toth et al., 2022) 87 Surface 0,1 (Toth et al., 2022) 88 N.D. N.D1,32 Iran Surface - (Pakzad et al., μg/L water 2023) 89 NI NI-0,011ppb Iran Groundwater - (Samarghandi et al., 2023) 89 NI NI-0,011ppb Iran Groundwater - (Samarghandi et al., 2023) Surface 0,1 (Szekacs; Moertl; water 2023) 0,05-3μg/L NI-0,011ppb Iran Groundwater - (Samarghandi et al., 2023) 0,05-3μg/L NI-0,011ppb Iran Groundwater - (Samarghandi et al., 2023) 0,05-3μg/L NI-0,011ppb Iran Groundwater - (Samarghandi et al., 2023) 0,05-3μg/L NI-0,011ppb Iran Groundwater - (Samarghandi et al., 2023) 0,05-3μg/L NI-0,011ppb Iran Groundwater - (Samarghandi et al., 2023)		Republic)					
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80 0,1- 0,1-47μg/L France 386,9μg/L Surface water 0,1 0,1-47μg/L (Lefrancq et al., 2017) 81 367- 1100ng/L 300- 1500ng/L France water Surface Surface 0,1 0,40- 12,9μg/L (Reoyo-Prats et al., 2017) 82 3,66- 12,9μg/L 0,40- 1,19μg/L France water Surface water 0,1 0,02-130μg/L (Desmet et al., 2025) 83 Netherland s Surface water 0,1 2016) (Desmet et al., 2016) 0,02-130μg/L <0,02-12μg/L Netherland s Surface water 0,1 2020) (Geerdink et al., 2020) 85 N.D. N.D. Netherland s Groundwater 0,1 3 (Broers et al., 2024) 86 Surface 0,05-3μg/L 0,2-2μg/L Hungary yeater Water 0,1 3 (Szekacs; Moertl; Darvas, 2015) 87 Suface 0,1 3 0,1 3 (Szekacs; Moertl; Darvas, 2015) 88 N.D. N.D1,32 μg/L Iran water Surface 900undwater - 4 (Pakzad et al., 2023) 89 NI NI-0,011ppb Iran Iran Groundwater - 5 (Samarghandi et <th></th> <th></th> <th></th> <th>•</th> <th></th> <th></th> <th></th>				•			
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Netherland Surface O,1 (Geerdink et al., 2016)							
81 367- 1100ng/L 300- 1500ng/L France water Surface water 0,1 2025) (Reoyo-Prats et al., 2017) 82 3,66- 12,9µg/L 0,40- 1,19µg/L France water Surface water 0,1 2025) (Merdy et al., 2025) 83 Netherland s Surface water 0,1 2016) (Desmet et al., 2016) 84 111- 9900ng/L Netherland s Surface water 0,1 2020) (Geerdink et al., 2020) 85 N.D. N.D. Netherland s Groundwater 0,1 2024) (Broers et al., 2024) 86 Surface 0,05-3µg/L 0,2-2µg/L Hungary Hungary water 0,1 3024 (Szekacs; Moertl; 2022) N.M. Surface water; 0,1 25-455ng/L (Szekacs; Moertl; 2023) Darvas, 2015) 88 N.D. N.D1,32 µg/L Iran yater Surface yater - (Pakzad et al., 2023) 89 NI NI-0,011ppb Iran yater Groundwater - (Samarghandi et	80	0.4.47//	•	France		0,1	
1100ng/L 1500ng/L water al., 2017	-04					0.4	,
82 3,66- 12,9μg/L 0,40- 1,19μg/L France water Surface yater 0,1 2025 (Merdy et al., 2025) 83 Netherland s Surface water 0,1 2016) (Desmet et al., 2016) 84 111- 9900ng/L Netherland s Surface water 0,1 2020) (Geerdink et al., 2020) 85 N.D. N.D. Netherland s Groundwater 0,1 30,000 (Broers et al., 2024) 86 Surface 0,05-3μg/L 0,2-2μg/L Hungary Water 87 Suface water; 0,1 30,000 (Szekacs; Moertl; Darvas, 2015) 88 N.D. N.D1,32 μg/L Iran yater Surface water - 30,000 (Pakzad et al., 2023) 89 NI NI-0,011ppb Iran Iran Groundwater - 30,000 (Samarghandi et	81			France		0, 1	` •
12,9μg/L 1,19μg/L water 2025	82			France		0.1	- , - ,
83 Netherland s Surface water 0,1 (Desmet et al., 2016) 0,02-130μg/L 0,02-12μg/L <0,02-12μg/L Netherland Surface water 0,1 (Geerdink et al., 2020) 84 111- 9900ng/L 14-340 ng/L s water N.D. Netherland Surface water 0,1 (Broers et al., 2020) 85 N.D. N.D. N.D. Netherland s Surface N.D. Surface N.D. Surface N.D. Surface Water 0,1 (Toth et al., 2022) 86 Surface N.D. Surface Water 0,1 (Szekacs; Moertl; Darvas, 2015) N.M. Water; Darvas, 2015) Forundwater Surface N.D. N.D1,32 Iran Surface N.D. Water - (Pakzad et al., 2023) 88 N.D. N.D1,32 Iran Surface Number N.D. N.D1,32 Iran Surface Number N.D. Water - (Samarghandi et N.D. Surface N.D. N.D1,32 Iran Surface Number N.D. N.D1,32 Iran Surface Number N.D. N.D. N.D1,32 Iran N.D. N.D1,32 Iran N.D. N.D1,32 Iran N.D. N.D1,32 Iran N.D1,32	02		•	Trance		0, 1	
S Water 2016 S Water 2016 S Water	83	12,0µg/L	1,10µg/L	Netherland		0.1	
Netherland Surface O,1 (Geerdink et al., 2020)						0,1	
84 111- 9900ng/L Netherland 14-340 ng/L Netherland s Surface Water 0,1 2020) (Geerdink et al., 2020) 85 N.D. N.D. Netherland s Groundwater 0,1 2024) (Broers et al., 2024) 86 Surface 0,05-3μg/L 0,1 40,05-3μg/L (Toth et al., 2022) Year 87 Suface Water; N.M. 0,1 2023) (Szekacs; Moertl; Darvas, 2015) 125-455ng/L Hungary Groundwater - 40,05-3μg/L (Pakzad et al., 2023) 88 N.D. N.D1,32 μg/L Iran Surface Water - 40,000 (Pakzad et al., 2023) 89 NI NI-0,011ppb Iran Groundwater - (Samarghandi et		0.02-130ua/L	<0.02-12ua/L	_			=0.07
9900ng/L 14-340 ng/L s water 2020	84		, ip:3:-	Netherland	Surface	0,1	(Geerdink et al.,
85 N.D. N.D. Netherland s Groundwater Groundwater 0,1 (Broers et al., 2024) 86 Surface 0,05-3μg/L 0,2-2μg/L Hungary Water 0,1 (Toth et al., 2022) 87 Suface Water; Suface Water; Darvas, 2015) N.M. Water; Groundwater 88 N.D. N.D1,32 Iran Surface Water - (Pakzad et al., 2023) 89 NI NI-0,011ppb Iran Groundwater - (Samarghandi et		9900ng/L	14-340 ng/L			,	
Surface O,05-3μg/L O,2-2μg/L Hungary Water	85			Netherland		0,1	,
N.M. Suface O,1 (Szekacs; Moertl; water; Darvas, 2015)				S			
87 Suface water; 0,1 Darvas, 2015) (Szekacs; Moertl; Darvas, 2015) 88 N.D. N.D1,32 Iran Surface μg/L water - (Pakzad et al., 2023) 89 NI NI-0,011ppb Iran Groundwater - (Samarghandi et	86				Surface	0,1	(Toth et al., 2022)
N.M. water; Darvas, 2015) 125-455ng/L Hungary Groundwater 88 N.D. N.D1,32 Iran Surface - (Pakzad et al., yater) μg/L water 2023) 89 NI NI-0,011ppb Iran Groundwater - (Samarghandi et		0,05-3µg/L	0,2-2µg/L	Hungary	water		
125-455ng/L Hungary Groundwater 88 N.D. N.D1,32 Iran Surface - (Pakzad et al., pate) µg/L water 2023) 89 NI NI-0,011ppb Iran Groundwater - (Samarghandi et al., pate)	87					0,1	
88 N.D. N.D1,32 μg/L Iran surface water - (Pakzad et al., 2023) 89 NI NI-0,011ppb Iran Groundwater - (Samarghandi et al.)		N.M.			water;		Darvas, 2015)
μg/L water 2023) 89 NI NI-0,011ppb Iran Groundwater - (Samarghandi <i>et</i>							
89 NI NI-0,011ppb Iran Groundwater - (Samarghandi et	88	N.D.		Iran		-	
, 11							
<i>al.</i> , 2017)	89	NI	NI-0,011ppb	Iran	Groundwater	-	
							al., 2017)



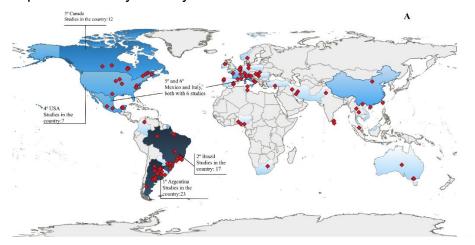
90				Surface	-	(Kalantary;
	N.M.			water		Barzegar; Jorfi,
		0,05-10µg/L	Iran			2022)
91	N.M.	0,01-0,04µg		Surface	_	(Farrow <i>et al.</i> ,
٠.	14.171.	I	Ireland	water		2025)
92	0,3-30,2µg/L	0,026-	Italy	Surface	0,1	(Campanale <i>et</i>
92	0,3-30,2µg/L	•	пату		0, 1	
		0,55µg/L		water		al., 2024)
93	N.D.	95-260ng/L	Italy	Surface	0,1	(Feltracco et al.,
				water		2022a)
94	<50-	<50-	Italy	Groundwater	0,1	(Suciu <i>et al.</i> ,
	8500ng/L	5500ng/L	•			2023)
95	<u></u>		Italy	Surface	0,1	(Masiol; Giannì;
00	0.05.1.4ug/l	0.06.2.1ug/l	italy	water	0,1	•
	0,05-1,4µg/L	0,06-2,1µg/L	14 - 1		0.4	Prete, 2018)
96	N.M.	0,17-	Italy	Surface	0,1	(Centanni <i>et al.</i> ,
		0,42µg/L		water		2024)
97	N.M.			Surface	-	(Joni et al., 2021)
		2,33-200µg/L	Malaysia	water		
98	N.D.	0,87-	Mexico	Surface	10	(Reynoso <i>et al.</i> ,
		4,33µg/L		water		2020)
99	1,74-7,4µg/L	N.D8,32	Mexico	Groundwater	10	(Osten <i>et al.</i> ,
33	1,14-1,4µg/L		INICKICO	Groundwater	10	
		µg/L				2025)
100	N.D.	0,35-1,4µg/L	Mexico	Suface	10	(Osten; Dzul-
				water;		Caamal, 2017)
				Groundwater		
101	N.M.	0,44-	Mexico	Groundwater	10	(Osten; Dzul-
-		1,42µg/L				Caamal, 2017)
102	N.M.	56,96-	Mexico	Surface	10	(Silva-Madera <i>et</i>
102	IN.IVI.		MEXICO	water	10	
400		510,6ppb	N 4		40	al., 2021)
103			Mexico	Surface	10	(de J. Bastidas-
	0,413-	<lod< th=""><th></th><th>water</th><th></th><th>Bastidas <i>et al.</i>,</th></lod<>		water		Bastidas <i>et al.</i> ,
	0,472µg/L					2024)
104	N.M.	N.M		Surface	-	(Nwinyimagu et
		0,33µg/L	Nigeria	water		al. 2023)
105	N.M.	N.D	J	Surface	_	(Tongo <i>et al.</i> ,
		0,00002µg/L	Nigeria	water		2022)
106		0,00002μg/L	rtigoria	Suface		(Ayoola <i>et al.</i> ,
100	NI NA	4.00			-	
	N.M.	1,02-		water;		2023)
		22,54µg/L	Nigeria	Groundwater		
107	N.M.			Surface	0,1	(Stenrod, 2015)
		0,17-12µg/L	Norway	water		
108	N.M.	N.D		Surface	-	(Khan <i>et al.</i> ,
		0,4593ppm	Pakistan	water		2020)
109		о, госоррии		Surface	_	(Urbańska-
100	0,131-	N.M.		water		Kozłowska et
		IN.IVI.	Dolond	water		
440	0,2411µg/L	0.00.4.00	Poland	0	0.4	al.2025)
110	0,05-4,24	0,03-4,69	Portugal	Groundwater	0,1 µg/L	(Inês; Ana; Silva,
	μg/L	μg/L				2024)
111	1,019-	<lod< th=""><th></th><th>Surface</th><th>0,1</th><th>(Vlassa <i>et al.</i>,</th></lod<>		Surface	0,1	(Vlassa <i>et al.</i> ,
	7,621µg/L		Romania	water		2022)
112	0,029-	0,005-		Surface	0,1	(Agarski <i>et al.</i> ,
	2.006µg/L	0,503µg/L	Serbia	water	,	2024)
113	2-11µg/L	1-45µg/L	Sri Lanka	Suface	_	(Gunarathna <i>et</i>
113	2 11µg/L	1 TOMB/L	On Lanka		_	•
				water;		<i>al.</i> , 2018)
				Groundwater		
114				Suface	-	(Jayasumana <i>et</i>
	N.M.			water;		<i>al.</i> , 2015)
		0,05-3,5µg/L	Sri Lanka	Groundwater		
_						



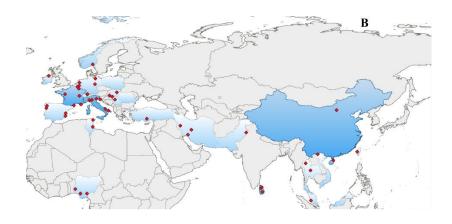
115	N.M.	0,36-		Groundwater	-	(Ulrich <i>et al.</i> ,
		7,31µg/L	Sri Lanka			2023)
116	N.M.	N.D.		Surface		(Jayasiri <i>et al.</i> ,
			Sri Lanka	water		2022)
117	N.M.	0,047-		Surface	0,1	(Poiger <i>et al.</i> ,
		8,7µg/L	Switzerland	water		2020)
118		N.D	Switzerland	Surface	0,1	(Poiger <i>et al.</i> ,
	N.D1,3µg/L	0,38µg/L		water		2017)
119			Switzerland	Surface	0,1	(Huntscha et al.,
	24-1680ng/L	<5-1430ng/L		water		2018)
120	0,04-	0,04-0,1 µg/L	Thailand	Surface	-	(Klaimala <i>et al.</i> ,
	4,34µg/L			water		2022)
121		Not		Surface	-	(Patel <i>et al.</i> ,
	Not available	available-		water		2024)
		0,47µg/L	Thailand			
122	N.D1509	N.D1354	Taiwan	Surface	-	(Lin; Chang;
	ng/L	ng/L		water		Sheen, 2022)
123	N.D.	0,1-1,73 μg/L	Tunisia	Surface	-	(Dahmeni et
				water		al.2024)
124	0,610-76,70	0,86-0,490	Tunisia	Surface	-	(Grünberger <i>et</i>
	μg/L	μg/L		water		al., 2024)
125	N.M.	0,88-		Surface	-	(Aydin <i>et al.</i> ,
		4,64ppm	Turkey	water		2023)
126				Surface	-	(Rodríguez-
		227,5-		water		Bolaña <i>et al.</i> ,
	580-890ng/L	1045ng/L	Uruguay			2023)
127			North	Surface	-	(Vu <i>et al.</i> , 2023)
	6,9-1329ng/L	54,5-565ng/L	Vietnam	water		

(N.D): Not detected; (N.M): Not measured; Groudwater: well water and deep groundwater; Surface water: river water, lakes, wastewater, tap water, and streams; -: No corresponding value found; *Research conducted in various European countries. Source: Prepared by the authors.

Figure 2Distribution of publications by country from 2015 to 2025.







3.1.2 Detected concentrations of glyphosate and AMPA and their regulatory limits

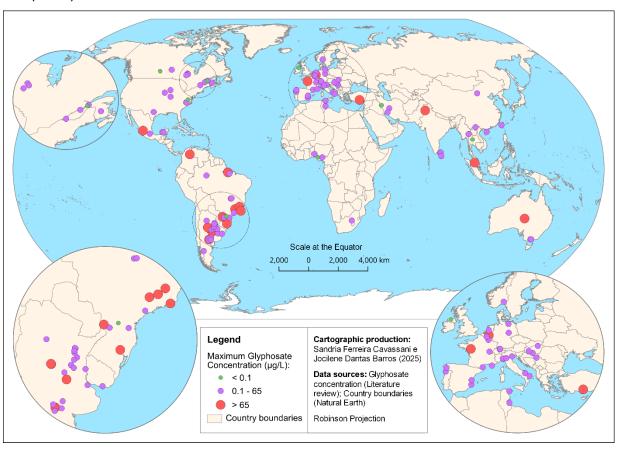
Fig. 3 depicts the global distribution of maximum glyphosate concentrations reported. Statistical analysis (Kruskal–Wallis test) revealed a significant location effect on detected concentrations [$x^2(2) = 15.079$; p < 0.001], with post-hoc results indicating significant differences particularly between Europe and North America (p = 0.736). Overall, concentrations were consistently higher in South and North America, with Brazil, Argentina, Colombia, and Mexico recording the most critical values. To facilitate interpretation, concentrations were categorized relative to prevailing drinking water regulations: (a) \leq 0.1 μ g/L (European Union standard, most stringent), (b) 0.1–65 μ g/L (range encompassing Brazilian and North American limits), and (c) >65 μ g/L (above most national standards) (Table 1). The mapping analysis showed that the vast majority of records worldwide exceeded the EU limit (0.1 μ g/L), including in European sites where lower concentrations would be expected under stricter regulation.

South America emerged as the most impacted region, with a high density of sites exceeding 65 μ g/L (red markers in Fig. 3). Brazil presented the highest glyphosate concentration globally (8,700 μ g/L;Lima *et al.*, 2023) — a value 133 times the national permissible limit (65 μ g/L) and approximately 87,000 times higher than the EU guideline. Argentina recorded a maximum of 1,600 μ g/L (Avigliano; Schenone, 2015b), while Mexico (510.6 μ g/L Silva-Madera *et al.*, 2021) and Colombia (441,5 μ g/L;Alvarez Bayona *et al.*, 2022) also reported values well above their national or international thresholds. By contrast, Canada



presented the two lowest concentrations globally (0.00007 μ g/L and 0.0028 μ g/L; Picard et al., 2021; Montiel-León et al., 2019), comparable to ultra-trace levels.

Figure 3World map of the maximum glyphosate concentrations found. The highest concentrations were found in South America, in Brazil (mean: 3,214 μg/L) and Argentina (1,600 μg/L) with a p-value of (.736).



Source: Cavassani; Barros, 2025 and research database.

For AMPA, the highest concentrations were likewise documented in Brazil and Argentina (de Matos *et al.*, 2023; Vera-Candioti *et al.*, 2021a) and in Belgium (Quaglia *et al.*, 2024). confirming a pattern of geographical overlap between glyphosate use intensity and metabolite accumulation (**Table 1**). Conversely, the lowest AMPA values were reported in Switzerland, Canada, and Vietnam (Picard *et al.*, 2021; Vu *et al.*, 2023).

A comparison with existing global pesticide reviews (De Araújo et al., 2022; Souza et al., 2020) highlights a critical gap: these earlier studies did not include glyphosate or AMPA, despite glyphosate being the most widely used herbicide worldwide. This omission underscores the novel contribution of the present review, which integrates both systematic



and scientometric approaches to demonstrate that glyphosate remains underrepresented in quantitative aquatic monitoring studies Pires *et al.* (2023). The mismatch between widespread application and limited monitoring points to significant environmental surveillance gaps.

Regarding legislation, most countries included in the 35-study dataset have established limits for glyphosate, including all EU nations, the US, Canada, Mexico, Brazil, Argentina, Paraguay, and Uruguay. However, many Asian countries (e.g., Malaysia, Sri Lanka, Thailand, Taiwan, Pakistan) and South Africa lack specific regulations, often defaulting to EU standards (as in China). Among countries with formal thresholds, the EU has the most restrictive value (0.1 μ g/L), while Australia allows the highest (1,000 μ g/L), followed by the United States (700 μ g/L, groundwater), Brazil (500 μ g/L, groundwater), Canada (280 μ g/L), and Argentina (240 μ g/L).

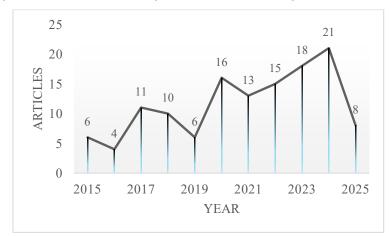
4 SCIENTOMETRIC DATA ANALYSIS

4.1 PUBLICATIONS AND CITATION NETWORK OVER THE YEARS

From 2020 onward, the number of publications addressing glyphosate concentrations in aquatic systems showed a marked increase, peaking in 2024 with 21 articles (Fig. 4). This surge in research output is likely associated with the intensification of debates over environmental and public health risks and the growing international concern regarding water resource contamination. Moreover, the development and dissemination of sensitive and specific analytical methods in recent years appear to have facilitated the expansion of monitoring studies.



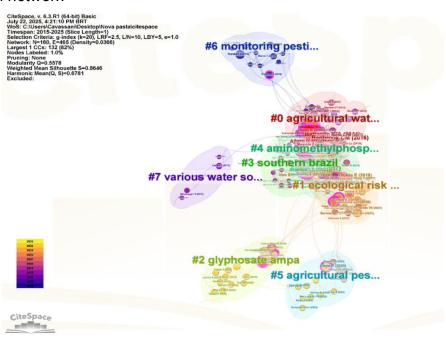
Figure 4
Number os article publications over the year. Source:Survey Data



To examine publication influence and author prominence, CiteSpace was employed to construct co-citation and collaboration networks. In **Fig. 5**, nodes highlighted in pink indicate citation bursts, where sudden increases in citation frequency reflect elevated visibility and influence of specific authors across the 127-article corpus. These bursts provide insight into researchers who have gained intellectual prominence within this field.

Figure 5

Author Citation network



Source: CiteSpace.



The network structure revealed eight distinct thematic clusters (#0–#7), with strong modularity (Q = 0.5578) and high silhouette values (S = 0.8646), confirming robust cluster separation and coherence. The largest cluster (#0 Agricultural Watershed) encompassed 30 studies, anchored by Okada *et al.* (2018), who investigated point-source glyphosate and AMPA pollution in Argentina. The second largest (#1 Ecological Risk Assessment) grouped 24 studies with a focus on ecotoxicological frameworks, led by Vera-Candioti *et al.* (2021a). Specialized research domains were identified in Clusters #2 (Glyphosate–AMPA) and #4 (Aminomethylphosphonic Acid), addressing degradation pathways and metabolite dynamics. Notable studies include Pires *et al.* (2023a), analyzing watershed contamination in Brazil, and Medalie *et al.* (2020), monitoring glyphosate residues in U.S. streams. The remaining smaller clusters (#5–#7) focused on pesticide monitoring across diverse aquatic matrices.

Citation analysis further identified the 10 most highly cited studies (**Table 2**) Bradley et al. (2017) emerged as the most influential, with 255 citations for the article Expanded Target-Chemical Analysis Reveals Extensive Mixed-Organic-Contaminant Exposure in U.S. Streams. Strikingly, half of the top ten most cited papers were published in 2017, highlighting this year as a pivotal milestone in glyphosate research. This temporal concentration coincides with the period following IARC's 2015 classification of glyphosate as a probable carcinogen and the subsequent FAO/WHO re-evaluation in 2016, events that intensified scientific and regulatory discourse globally. The dissemination of advanced analytical methodologies during this period further catalyzed research expansion, reinforcing the interplay between regulatory controversy, methodological innovation, and knowledge production.

In **Fig. 5**, small cluster density are observed, which indicate an increase in thematic publications concerning glyphosate and AMPA concentration in water courses. Network analysis in CiteSpace revealed 8 thematic clusters (ranging from #0 to #7) that represent the principal research foci on glyphosate and AMPA contamination in water courses. The high modularity represented by Q= 0.5578 and the silhouette value (S= 0.8646) indicate that the clusters are well-defined and coherent. The most dominant cluster (#0 agricultural watershed) possesses 30 members and has as its central article the study by Okada et al. (2018), published in Environmental Science and Pollution Research, on point pollution of glyphosate and AMPA in a rural watershed in Argentina. The second largest cluster (#1 ecological risk assessment) possesses thematic focus directed toward ecotoxicological approaches and contains 24 members. The central article is the study by Vera-Candioti et al. (2021a),



published in Environmental Monitoring and Assessment, on environmental monitoring. Clusters #2 glyphosate AMPA and #4 aminomethylphosphonic acid are thematic areas focused on glyphosate degradation and its metabolite AMPA. Within these two thematic areas, notable authors include Pires et al. (2023a) with the study conducted in a watershed in Brazil (as well as #3 southern Brazil) and Medalie et al. (2020), a study conducted in streams in the United States. Clusters #5, #6, and #7 are formed by small members and are focused on the thematic area of pesticide monitoring in various water sources.

Regarding authors and publications, **Table 2** obtained from WoS delineated the 10 articles (within the 127 articles) most cited in studies on glyphosate and AMPA concentrations. The table also shows the values of total and annual citation counts for each article, in addition to information about the authors and the year in which the article was published. Through citation analysis of the articles, it can be perceived that the most influential author is Bradley et al. (2017), with a total of 255 citations in the article entitled "Expanded Target-Chemical Analysis Reveals Extensive Mixed-Organic-Contaminant Exposure in U.S. Streams." It is also possible to note that of the 10 most cited articles, 5 (five) were published in the year 2017, indicating that this period represented a relevant milestone for the advancement of research on the topic, and may be directly related to the international regulatory context subsequent to the classification of glyphosate by IARC in 2015 and the review by FAO and WHO in 2016, which intensified the global scientific debate regarding the effects of glyphosate exposure. Additionally, the dissemination of more sensitive analytical methodologies may have propelled advancement in research.

 Table 2

 10 most cited articles on glyphosate and AMPA concentrations in watercourses

Author, year	Title	Total	Average
		citations	per year
Bradley et al.,	Expanded Target-Chemical Analysis Reveals	255	28,33
2017	Extensive Mixed-Organic-Contaminant Exposure		
	in U.S. Streams		
Primost et al.,	Glyphosate and AMPA, "pseudo-persistent"	187	20,78
2017	pollutants under real-world agricultural		
	management practices in the Mesopotamic		
	Pampas agroecosystem, Argentina		



122	М٠	2358-	-2472
133	•	2000	Z4/Z

Jayasumana <i>et</i>	Drinking well water and occupational exposure to	155	14,09
al., 2015	Herbicides is associated with chronic kidney		
	disease, in Padavi-Sripura, Sri Lanka		
Ronco et al.,	Water quality of the main tributaries of the Paraná	132	13,20
2016	Basin: glyphosate and AMPA in surface water and		
	bottom sediments		
Poiger et al.,	Occurrence of the herbicide glyphosate and its	126	14,00
2017	metabolite AMPA in surface waters in Switzerland		
	determined with on-line solid phase extraction LC-		
	MS/MS		
Lefrancq et al.,	High frequency monitoring of pesticides in runoff	116	12,89
2017	water to improve understanding of their transport		
	and environmental impacts		
Lupi <i>et al.</i> , 2015	Occurrence of glyphosate and AMPA in an	114	10,36
	agricultural watershed from the southeastern		
	region of Argentina		
(Alonso et al.,	Glyphosate and atrazine in rainfall and soils in	111	13,88
2018)	agroproductive areas of the pampas region in		
	Argentina		
(Mahler et al.,	Similarities and differences in occurrence and	97	10,78
2017)	temporal fluctuations in glyphosate and atrazine in		
	small Midwestern streams (USA) during the 2013		
	growing season		
(Demonte et al.,	Determination of glyphosate, AMPA and	90	11,25
2018)	glufosinate in dairy farm water from Argentina		
	using a simplified UHPLC-MS/MS method		
e: Research Data		l	<u> </u>

Source: Research Data.

4.1.1 Most relevant keywords and cluster analysis

Keyword co-occurrence analysis was performed in CiteSpace, generating a structured network composed of 257 nodes and 1,022 links (**Fig. 6**). The network presented acceptable modularity (Q = 0.3587) and satisfactory internal consistency (S = 0.5121), indicating that the identified clusters are thematically coherent and represent a well-organized knowledge domain.

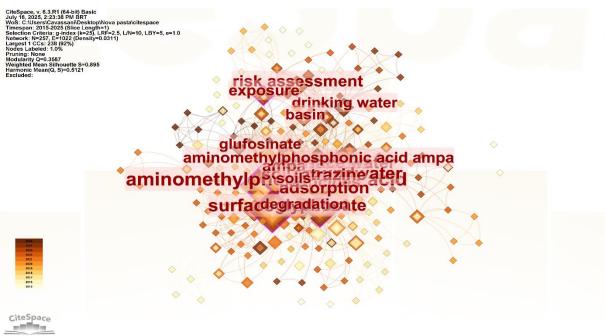
The most frequent keyword was "surface water" (45 co-occurrences), highlighting the predominance of studies focused on rivers and lakes adjacent to agricultural areas. This was followed by "aminomethylphosphonic acid" (36), reflecting the growing research attention given to glyphosate's primary metabolite (AMPA). The keyword "glyphosate" (26) indicated



the parent compound as a central axis of investigation, while "soil" (23) suggested frequent consideration of soil-to-water leaching pathways. Together, these terms delineate the ecological interface between agricultural management, environmental transport, and freshwater contamination.

Figure 6

Most used Keywords and Citation bursts



Temporal burst analysis provided further insights into evolving research priorities (Fig. 7). The keyword "streams" exhibited the strongest burst (intensity 3.81) between 2015–2017, corresponding to early investigations of contaminant transport in watershed systems. The methodological term "solid phase extraction" showed a burst strength of 3.68, with peak activity in 2021, reflecting advances in extraction and detection methods that have substantially increased monitoring sensitivity for glyphosate and AMPA. This technical evolution coincided with the surge in publication output observed after 2021 (Fig. 2), suggesting a clear link between methodological innovation and research productivity.

In addition, the keyword "pesticides" (burst 3.41, peak in 2018) revealed a broadening of research scope toward comparative assessments across multiple agrochemicals, while "environmental fate" (burst 3.39) and "impact" (burst 3.25) indicate a growing emphasis on ecological persistence and human health risks. These terms collectively signal a transition from studies limited to chemical occurrence toward more integrated approaches addressing



exposure pathways, toxicological outcomes, and regulatory implications. In summary, the keyword analysis highlights both the consolidation of glyphosate/AMPA research within traditional monitoring domains (surface waters, leaching) and the emergence of methodological and thematic innovations. The co-occurrence of environmental and toxicological terms illustrates a research field increasingly oriented toward understanding the ecological fate and health impacts of glyphosate, though notable geographic and thematic gaps remain.

Figure 7

Top 5 Keywords with the Strongest Ciation Bursts

Top 5 Keywords with the Strongest Citation Bursts

Keywords	Year	Strength	Begin	End	2015 - 2025
streams	2015	3.81	2015	2017	
solid phase extraction	2021	3.68	2021	2021	
pesticides	2016	3.41	2018	2018	
environmental fate	2017	3.39	2017	2018	
impact	2019	3.25	2024	2025	

Source: Cite Space.

4.1.2 Countries and cooperation analysis

Country-level analysis of research productivity, based on authors' institutional affiliations, provides critical insight into the geographic distribution of studies on glyphosate and AMPA concentrations in aquatic systems. As shown in **Fig. 8**, the cooperation network reveals both publication output and international collaboration clusters, with Argentina emerging as the most productive nation, leading with 26 publications, followed by Brazil (18), the United States (15), Canada (12), France (9), Italy (7), Switzerland (6), Spain (6), the Netherlands (5), and Mexico (5).

Beyond publication counts, betweenness centrality was employed to assess the influence each country exerts within the global research network (Chen, 2014). Argentina and the United States exhibited the highest centrality (0.29 each), followed by France (0.20), Italy (0.16), the Netherlands (0.09), Brazil and Mexico (0.06 each), Switzerland (0.05), and Spain (0.01). These results demonstrate that while Argentina and Brazil dominate in absolute

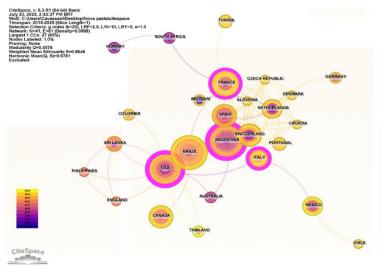


productivity, the network of international intellectual influence is more strongly anchored in the United States and several European countries.

This divergence highlights a structural asymmetry: in Argentina, high centrality appears linked to its dependence on transgenic soybean agribusiness, where intensive glyphosate use drives both environmental monitoring and international visibility of related studies (Fernandes *et al.*, 2019), In contrast, the United States' elevated centrality derives not only from research volume but also from its ability to mobilize large-scale collaborative networks supported by institutional funding, which positions it strategically in shaping international scientific and regulatory agendas. The impact of this position is exemplified by Bradley *et al.* (2017), the most highly cited study in this field. This contrast—between major producer countries in Latin America (Argentina and Brazil) and countries with stronger collaborative infrastructure (United States and Europe)—underscores a persistent imbalance between research productivity and global influence. Addressing this gap will require strengthening regional scientific cooperation in Latin America and reducing dependency on hegemonic research centers, thereby promoting more equitable participation in the generation of knowledge and in the formation of international environmental policies.

Figure 8

Collaboration network among the countries.



Source: CiteSpace.

5 CONCLUSION

This review demonstrates that the highest glyphosate concentrations occur predominantly in South and North America, with Brazil reporting extreme values that surpass



the national regulatory threshold by more than two orders of magnitude. Scientometric analysis revealed a consistent upward trend in publications, reflecting heightened environmental concern and increasing toxicological evidence linking glyphosate and its metabolite AMPA to adverse human health outcomes. Despite this progress, important gaps remain: knowledge of the environmental pathways and long-term health impacts of these compounds is still fragmented, and even countries with stringent regulatory frameworks frequently report concentrations that exceed established limits. Such findings expose persistent weaknesses in monitoring and enforcement mechanisms. Taken together, the results highlight an urgent need to strengthen and harmonize global monitoring programs, foster greater international cooperation—particularly in underrepresented regions such as Asia and Africa—and integrate scientometric insights with environmental surveillance to identify emerging research fronts and inform more effective regulatory policies. By uniting evidence of widespread contamination with an evaluation of global research patterns, this study underscores glyphosate's dual role as an environmental pollutant and a matter of international scientific and political relevance, demanding coordinated action to protect both ecosystems and human health.

Future research should focus on expanding monitoring efforts in underrepresented regions, developing standardized analytical methods, and establishing long-term monitoring programs to better understand temporal trends and the effectiveness of regulatory interventions.

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DECLARATIONS

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